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Address of Editorial office:

Agricultural Science and Technology
Faculty of Agriculture, Trakia University
Student's campus, 6000 Stara Zagora
Bulgaria
Telephone: +359 42 699330
+359 42 699446
www.agriscitech.eu

Technical Assistance:

Nely Tsvetanova
Telephone: +359 42 699446
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Heavy metal pools in urban soils from city parks of Sofia, Bulgaria

V. G. Kachova^{1*}, I. D. Atanassova²

¹Forest Research Institute, Kliment Ohridski Blvd. 132, 1756 Sofia, Bulgaria

²Nikola Poushkarov Institute of Soil Science, Agrotechnology and Plant Protection, 1080 Sofia, Bulgaria

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Abstract. With a population ~ 1 300 000 million and an area of 1344 km², Sofia, Bulgaria is one of the most densely populated cities in Europe with registered contamination of the air, water and soils of the city including heavy metals. The purpose of the study is to analyze the different pools under which heavy metals in urban soils exist. The chosen sites are the largest forest parks in Sofia. Cu, Zn, Pb and Cd in soil of samples were sequentially extracted and separated soluble F1 (70% ethyl alcohol solution), liable F2 (1N CH₃COONa, pH = 7), carbonate bound F3 (1N CH₃COONa + 1N CH₃COOH, pH = 3), soil organic matter bound F4 (30% H₂O₂), amorphous iron oxide bound F5 (0.5 M Na-citrate), total "aqua regia" content and residual fraction F6 (measured as a difference from total). These operationally defined fractions give estimates of available (F1, F2), potentially liable (F3, F4, F5) and pseudo-total contents in soils, respectively. Metals were mainly concentrated in the "residual fraction" F6. Organic matter associated is mainly Cu and Pb, while carbonates-bound and amorphous iron oxides-bound were Pb and Zn. The mobility of the studied metals decreased in the order: Cd > Zn > Cu ≥ Pb. The soils with high pH and Cation exchange capacity measured at pH = 8.2 (CEC_{8.2}) show lower amounts of mobile (easily-available) forms of heavy metals.

Keywords: urban soils, heavy metals, sequential extraction

Introduction

Distinctive feature of soils from urban areas is the presence of a surface layer enriched with heavy metals (Kabata-Pendias and Pendias, 1992; Stroganova and Agarkova, 1993; Stroganova et al., 1998; Manta et al., 2002; Seleznev and Yamoshenko, 2014; Alekseenko and Alekseenko, 2014). Elevated concentrations of heavy metals (HM) in the top-soil negatively affect soil biota and plant biodiversity (Kapusta and Sobczyk, 2015) and conditions are created for disturbances in mineral nutrition, changes of metabolic and physiological processes in green plants. The deteriorated status of green systems directly reflects on the quality of the city environment. Soil pollution with heavy metals can have significant toxic effects on human health by inhalation of dust or direct contact in public green places, parks and gardens (Dolotov and Ponomareva, 1982; Obukhov and Kutukova, 1990). This fact determines the need of research concerning the content of HM in soils from parks of many cities around the world (Shao-Lian et al., 1989; Hou et al., 2015). The study of large parks in Stara Zagora (Bulgaria) reveals technogenic loads on soils with heavy metals on the bases of technogenic coefficient and coefficient of abnormality but without risks for ecosystems (Petkov et al., 2010). Links between soil pollution with heavy metals and age of the park are determined also (Li et al., 2001; Hou et al., 2015). Increased concentrations of Cd, Cu, Zn and Pb and other heavy metals were observed in the soil of the old city parks located in commercial and industrial areas, and this is attributed to the automobile and industrial activities in the areas (Tchuldjian, 2008; Qingjie et al., 2008; Du et al., 2013; Hussain et al., 2015). Determination of total content of metals is used in the identification and general description of soils, but to study their toxicity and behavior, distribution of heavy metals among different soil compartments is useful (Ladonin and Karpukhin, 2011; Minkina et al., 2014). By the term "pools of heavy metals" are not considered

* e-mail: vaniakachova@gmail.com

separate heavy metal compounds, but rather forms of compounds distinguished by many common features and properties (Ladonin, 2002). It is widely considered that the chemical forms of heavy metals determine their mobility and bioavailability to living organisms, and hence the strength of their toxicity (Ladonin, 2002; Lu et al., 2003; Doichinova et al., 2014). For determination of the mobile (bioavailable) forms of heavy metals are used either non-selective: NH₄OAc at pH 7; 2M HCl; NH₄OAc / EDTA / HCl: NH₄OAc / EDTA (Nowack et al., 2001; Mahava et al., 2003; Doichinova et al., 2014), or selective extractants for fractionating Cu in soils; for determination of non-silicate forms of Fe and Mn in soils (Mehra-Jackson, 1960; McLaren and Crawford, 1973). Over the years are used various successive sequential extraction treatments, where in the fractions are separated and defined as "pools of heavy metals in soils" (Lu et al., 2003; Janoš et al., 2010; Arenas-Lago et al., 2014).

The aim of the study is to analyze the different pools of heavy metals in urban soils from the largest city parks in Sofia.

Material and methods

Geographically, Sofia region is covering all Sofia valley with clearly differentiated kettle bottom, foothills and surrounding mountains with temperate climate. Frequent fogs are a characteristic feature. Temperature inversions, which determine the removal of ground air of polluting substances toward the higher levels, are also common phenomenon. Annual emissions of SO₂ and NO₂ are over 100 t / km² for 2010. These indicators determine the capital city as one of the most polluted towns in the country. Dust and metal aerosols are durable components in the air and meteorological conditions make it possible for their dissemination around the whole area.

In the soil surface (~ 20 cm) is located a large part of the fine

roots of trees. We conducted our research in the surface soil layers (0 – 20cm). The measured heavy metals are Cu, Zn, Pb and Cd, which are prevalent pollutants in urban environment (Doichinova et al., 2006). The study was carried out in the spring time month of April, 2014, in four parks of the capital, located in four directions from the centre:

U1 - Boris Garden at a distance from the city center 3.3 km, average sea level 550 m, slope 4, exposure: east-southeast, the predominant tree species: summer oak (*Quercus pedunculiflora* C. Koch) and soils *Anthrosols* with Hortic horizon (WRB, 2014), and deeper with characteristics of anthropogenic Vertisols. The morphological disruption of soil profile till parent rocks is: A'h (0-10cm); A" (10-30cm); A''' (30-53cm); AB (53-73cm); B (73-83cm); BC (83-110 ↓ cm). The total area of park is 3.02 km² from which was chosen an area of 0.05 ha for sample plot (SP) with homogenous vegetation.

U2 - Loven Park at a distance from the city center 3.7 km, average sea level 570 m, slope 4, exposure: south-southeast, predominant tree species: red oak (*Quercus rubra* L.) and soils *Anthrosols* with Hortic horizon (WRB, 2014), and deeper with characteristics of Haplic Luvisol. The morphological disruption of soil profile till parent rocks is: A'h (0-6cm); A" (6-34cm); B₁ (34-60cm); B₂ (60-80cm); BC (80-110 ↓ cm). The total area of park is about 0.5 km² from which was chosen an area of 0.05 ha for SP with homogenous vegetation.

U3 - West Park at a distance from the city center 3.8 km, average sea level 525 m, slope 6, exposure: west-southwest, predominant tree species: red oak (*Quercus rubra* L.) and soils *Anthrosols* with Hortic horizon (WRB, 2014) and deeper with characteristics of anthropogenic Vertisols. The morphological disruption of soil profile till parent rocks is: A'h (0-10cm); A" (10-30cm); A''' (30-73cm); AB (73-83cm); B (73-105cm); BC (105-110 ↓ cm). The total area of park is about 2.16 km² from which was chosen an area of 0.05 ha for SP with homogenous vegetation.

U4 - North Park at a distance from the city center 5.6 km, average sea level 575 m, slope 4, exposure: northeast predominant tree species red oak (*Quercus rubra* L.) and soils *Anthrosols* with Hortic horizon (WRB, 2014). They have characteristic of Haplic *Anthrosols* according to updated Bulgarian classification (Gencheva, 2000). The morphological disruption of soil profile till parent rocks is: I_{ah} (0-3cm); II (3-19cm); III (19-53cm); B (53-64cm); BC (64-83 ↓ cm). The total area of park is about 0.78 km² from which was chosen an area of 0.05 ha for SP with homogenous vegetation.

Soil samples (0-20cm) were taken at fourth replicates in the center of the plots plus one from soil profile (5 samples per plot and 20 samples in total). Soil characteristics were determined according to standardized methods (Donov et al. 1974) i.e., texture according to the method of Kachinsky (full description of all fractions is given by Doichinova et al. (2006), organic carbon (C%) according to the method of Turin, phosphorus – colorimetric method by Kirsanov (Donov et al., 1974).

The total content of heavy metals, macro and micro-elements in soil were measured after treatment with HCl: HNO₃ (ISO 11466) and analyzed by atomic absorption spectroscopy (AAS) on Perkin Elmer 370 A. Cation exchange capacity was determined by the method of Ganey (Ganey and Arsova, 1980). In order to determine the "forms of heavy metals in soils" we used the method of sequential extraction of Kuznetsov and Schimko (1990). Consistently, the following fractions were separated: F1 - "soluble pool" by 70% alcohol solution (to reduce the influence of ion exchange); F2 - "labile pool" by 1N CH₃COONa at pH 7; F3 "carbonates-bound pool" by a solution of 1N

CH₃COONa + 1N CH₃COOH at pH 3,5; F4 "organic matter- bound pool" by 30% H₂O₂; F5 "amorphous hydroxides bound pool" by 0.5 M Na-citrate. The fraction F6 "residual" or "insoluble" was calculated as a difference from total contents of heavy metals in soils (Dąvkowska-Naskręć, 2000). The ratio between the solid phase (soils) and the extraction solution is 1: 5.

In order to determine the heavy metal composition of fine roots (<= 2 mm) wet combustion by concentrated HNO₃, thermal ashing in a muffle furnace at 450°C, extraction with 20% HCl and analysis by AAS was used.

The data was present as mean and +/- SD; at significant level p < 0.05.

Results

General characteristic of soils (0-5; 5-10; 10-20 cm) and the average contents of heavy metals in fine roots of the tree species is given in Table 1.

Soils from the parks in the city of Sofia are slightly acidic to neutral with the exception of Loven Park where soils in depth are highly acidic. Textural composition varies from light sandy loam (Boris garden) to heavy sandy loam (Northern park). According to the values of CEC_{8.2} (cation exchange capacity at pH=8.2) soils are moderately colloidal (Boris garden and Loven park) to highly colloidal (Western and Northern park) and are characterized by high soil organic matter contents (Northern park). The content of HM in fine roots tends to decrease in the deeper layers. The average contents of heavy metals in the separate fractions from the 0-20cm layers the soils and the average total content (OF) are presented in Table 2.

The fraction of "soluble pool" (F1) has the lowest contents. The fraction of easily mobile "labile pool" (F2) represents predominantly ionic chemical bonds or weak complexes to soil components and is easily removed into soil solution. This fraction together with the fraction of the "soluble pool" (F1) is the most easily assimilated by plants. The fraction of the "carbonates-bound pool" (F3) is extracted with acidified solution of the extractant (pH = 3.5), and is most sensitive to a change in the acid status of the soil. Upon acidification these forms become available to the plants (Lu et al., 2003), and such acidification "in situ" may be caused, as a result of root exudates. In acid medium, the fraction F5 „amorphous Fe, Mn hydroxides bound" is also unstable, which defines this fraction as "potentially available" to plants. The fraction F4 "organic matter-bound" is potentially available for root extraction in case of alteration in redox conditions in soils and/or pH 5 and above (Alloway, 1995), which corresponded to most of our samples. Therefore, the total sum of the content of heavy metals in all these fractions, without fraction F6 "insoluble residue", represents the sum of the active (potentially digestible) metals in the soil (Xu and Yang, 1995).

Heavy metals in park soils of the city of Sofia exist mainly as "insoluble residue-pool" (F6). Active heavy metals are associated mainly with organic matter, carbonates and hydroxides of Fe and Mn. "Easily mobile pool" and "soluble pool" are not in large quantities.

Figures 1a, 1b, 1c, 1d represents a distribution of the heavy metals in different pools. In these figures are described separately Cu, Zn, Pb и Cd in sequentially extracted fractions: F1 - soluble; F2 - labile; F3 - carbonate-bound; F4 - organic matter-bound; F5 - amorphous hydroxides-bound; F6 - residual. In U1, U2, U3 and U4 from top towards bottom are represented layers: 0-5; 5-10; 10-20 cm

Table 1. Soil characteristics and average contents of heavy metals in fine roots (mean \pm 0.01)

Soil Classification Hortic Anthrosols (FAO)	pH (H ₂ O)	Silt + Clay %	CEC _{8.2} [*] cmol/kg	SOM ^{**} g/kg	P (mg/kg)	Cu in fine root mg/kg	Zn in fine root mg/kg	Pb in fine root mg/kg	Cd in fine root mg/kg
U1 Boris garden									
0-5 cm	6.65	26.54	30.04	45.0	292	16.33	44.34	26.97	1.07
5-10 cm	6.50	28.57	26.06	36.8	312	12.40	33.25	24.19	0.96
10-20 cm	6.60	32.72	27.01	27.4	336	13.81	35.04	23.20	0.82
U2 Loven park									
0-5 cm	5.10	20.53	28.15	53.7	410	14.39	31.86	20.24	0.63
5-10 cm	4.90	29.30	24.31	31.2	382	13.80	28.73	16.96	0.56
10-20 cm	4.50	32.27	27.92	21.7	332	12.24	29.28	16.59	0.50
U3 Western park									
0-5 cm	6.05	46.86	61.66	57.2	610	12.20	46.71	25.54	0.67
5-10 cm	5.80	48.39	65.42	39.3	590	12.62	38.01	20.03	0.63
10-20 cm	5.20	58.32	75.62	33.0	598	11.81	36.27	17.69	0.56
U4 Northern park									
0-5 cm	6.85	32.61	50.21	82.5	1139	19.05	35.77	17.54	0.52
5-10 cm	6.75	42.53	54.55	81.2	1023	21.39	30.32	14.99	0.59
10-20 cm	6.60	40.39	47.93	45.3	1069	18.03	27.79	13.02	0.45

n = 20

* CEC – cation exchange capacity in pH = 8.2

** SOM – soil organic matter

Table 2. Average contents of heavy metals fractions and average total contents (OF) in urban soils (mean \pm 0.1)

	F1	F2	F3	F4	F5	F6	OF
	mg/kg						
0-20 cm	Cu						
	0.24 \pm 0.0	0.54 \pm 0.0	1.57 \pm 0.2	5.79 \pm 0.4	0.87 \pm 0.1	36.27 \pm 7	45.27 \pm 7
0-20 cm	Zn						
	0.31 \pm 0.1	1.13 \pm 0.1	5.25 \pm 0.6	4.61 \pm 0.4	5.45 \pm 0.5	42.41 \pm 9.0	59.15 \pm 8.5
0-20 cm	Pb						
	0.10 \pm 0.1	0.91 \pm 0.1	1.93 \pm 0.4	2.04 \pm 0.3	1.68 \pm 0.2	22.42 \pm 1.3	29.08 \pm 1.5
0-20 cm	Cd						
	0.02 \pm 0.0	0.03 \pm 0.0	0.09 \pm 0.1	0.02 \pm 0.0	0.05 \pm 0.1	0.11 \pm 0.1	0.32 \pm 0.1

n = 20

*OF – total forms

The highest contents of heavy metals are in the "insoluble residue-pool". This pool is unreachable for plants, and is associated with the crystal lattice of soil minerals.

Generally, the distribution of metals in different fractions follows the order:

- Cu: F6 » F4 > F3 > F5 > F2 > F1
- Zn: F6 » F5 > F3 > F4 > F2 > F1
- Pb: F6 » F4 ~ F3 ~ F5 > F2 > F1
- Cd: F6 > F3 > F5 > F2 > F4 > F1

Copper and Pb are mainly associated with soil organic matter, while Zn and Cd with carbonates and Fe-Mn hydroxides.

Discussion

The analysis of heavy metal pools in soils is used for more detailed study of metals behavior in soils and for clarifying their

ability to migrate towards other components of the ecosystem. Determination of the static snapshot of soil structural components, with which metals are associated clarifies the mechanisms of interaction and behavior. Therefore, in many cases, especially when considering the availability of heavy metals and their assimilation by plants, it is important to determine their pools in soils. Heavy metal pools are exclusive prototype of geochemical diversity of different compounds of HM formed as a result of weathering processes (Tchuldzhiyan and Petrov, 1984). Separately, by anthropogenic import various compounds of heavy metals enter the soils and can be transformed depending on soil conditions. Changes in soil conditions (pH, % dust, etc.), as well as enhanced anthropogenic load influence the available pools of heavy metals and their transformations. Therefore, determination of heavy metal pools for particular soil conditions in urban parks is important for clarifying the behavior of metals, their accessibility, impact on mineral nutrition and assimilation by woody plants in urban environment.

The total concentration of the analyzed heavy metals do not

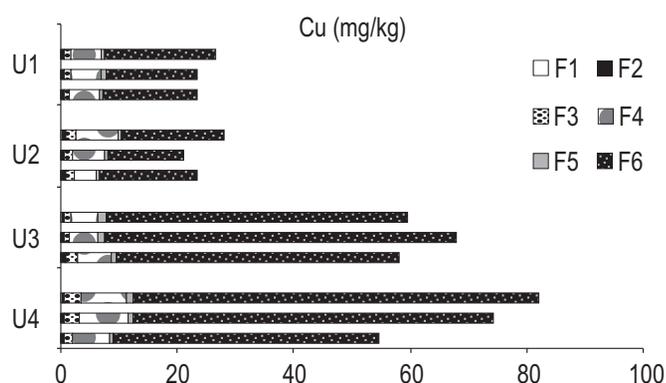


Figure 1a. Sequential fractions of Cu

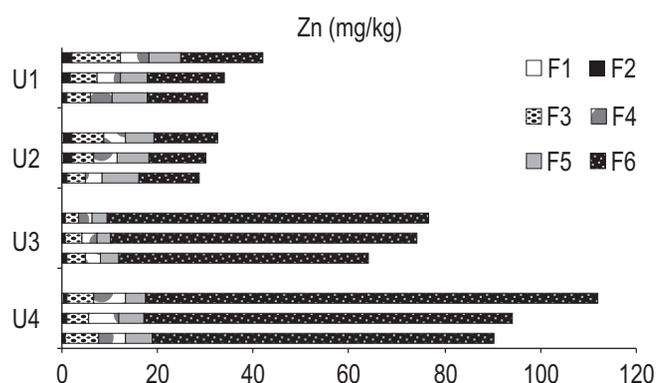


Figure 1b. Sequential fractions of Zn

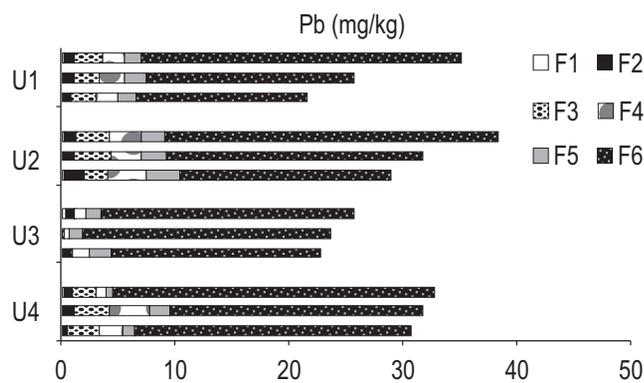


Figure 1c. Sequential fractions of Pb

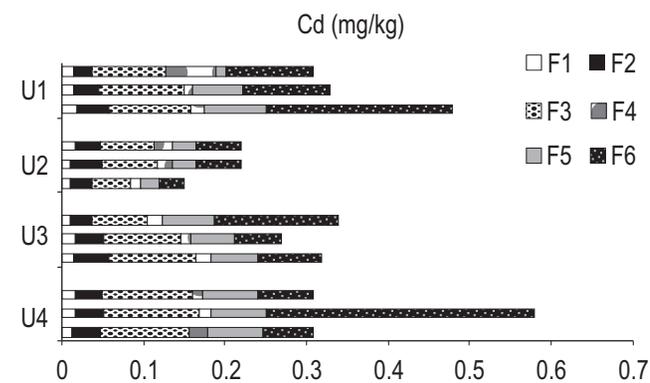


Figure 1d. Sequential fractions of Cd

exceed the official thresholds accepted in Bulgarian standards for soils from public areas (Regulation No. 3/1979; Regulation No. 3/2008).

The highest contents of Cu were in the residual fraction ($76 \pm 10\%$) and the lowest in the “soluble fraction” ($0.7 \pm 0.3\%$). The organically complexed Cu comprised $15.7 \pm 7\%$. The general Cu distribution was in the order: Total forms (TF) $\approx F6 \gg F4 > F3 > F5 > F2 > F1$. Active Cu is mainly related to soil organic matter (F4 – 16%). This affinity of Cu to soil organic matter was well defined by other authors (McLaren and Crawford, 1973; Alloway, 1995). Atanassova (1995) and Atanassova and Okazaki (1997) showed that at high copper levels, the clay fractions from acid soils could not play a scavenging role because Cu was exchangeable adsorbed by 2:1 clay minerals in soils.

Lead was primarily accumulated in the F6 residual fraction with substantial inputs in the organic fraction (F4 – 7%). Pb distribution followed the order: TF $> F6 \gg F4 \approx F3 \approx F5 > F2 > F1$. Daykovska-Naskręć (2000) found a high percentage of Pb bound to soil organic matter in urban soils of northern Poland. Lead shows high affinity to soil organic matter confirmed by previous studies in connection with the profile distribution of Pb in soils from the Sofia region (Doichinova and Atanassova, 2003). In the urban part of the region, Pb is also bound to carbonates (F3 - 6%) and Fe-Mn hydroxides (F5 - 6%). In F3 fraction are extracted metals from minerals such as: $Cu_2(OH)_2CO_3$, $Pb_3(PO_4)_2$, $Pb_5(PO_4)_3Cl$, $PbCO_3$, $Zn(OH)_6(CO_3)_2$, $ZnSiO_4$, hydroxides of heavy metals etc., which are present in soils as residual components from the weathering of parent rocks or as intermediates in the transformation of technogenic forms of HM (Alloway, 1995). Data on the distribution of Pb among different pools in urban soils support the results of the multiple analyses for anthropogenic load of soils in Sofia with Pb (Doichinova et al., 2006).

Zn also appeared in the residual fraction ($63 \pm 21.6\%$), but highest Zn pools were found in the carbonate and Fe-oxide fraction.

The overall trend followed TF $\approx F6 \gg F5 \approx F3 \approx F4 > F2 > F1$. Unlike Cu, the active Zn binds slightly less to the soil organic matter, rather than with Fe-Mn hydroxides (F5 – 13%) and carbonates (F3 – 11%). The same was found in studies of other authors concerning urban soils (Lu et al., 2003). Sedberry and Reddy (1976), however establish a higher percentage of organically bound Zn in agricultural soils of Louisiana, USA.

Compared with the other metals, Cd showed the least distribution in the residual fraction ($31 \pm 12.3\%$) and highest in the “labile” fraction ($11.4 \pm 3.9\%$), indicating potential toxic effects in the urban soils. Substantial proportions were also detected in the carbonate ($29.4 \pm 5.8\%$) and the Fe-hydroxide fractions ($16 \pm 5\%$).

Increased percentage of labile heavy metals is a characteristic feature of urban environment (Lu et al., 2003; Doichinova et al., 2014). Accessibility and mobility of HM in soils under oak ecosystems of urbanized soils from Sofia decreases in the following order: Cd $>$ Zn $>$ Cu $>$ Pb.

Cadmium is the most mobile and easily accessible element and Pb is the least accessible and has a pronounced accumulation in the top soil surface. The total loading of soils with Cd is results in increase of its active forms, and connected with the expression mostly of Cd-toxicity in the studied soils.

The relationships between various pools of heavy metals in different soils from the urban part (U) of the area and main soil characteristics are shown in Table 3.1, Table 3.2, Table 3.3 and Table 3.4.

Labile Cu (F2) is negatively correlated with pH and humus content. The content of clay is also negatively correlated with fractions F2, F3 and F5 of Zn which means that the surface of the soil of urban forest parks, does not render active clay fraction for adsorption of these Zn-fractions. It is interesting to note the positive correlation between organically bound Zn and pH. In the reported trend towards alkalinization in the surface soil horizons of urban (U)

Table 3. Correlation matrix of various heavy-metal fractions and soil characteristics (n=16)

Table 3/1. Cu

	F1	F2	F3	F4	F5	F6	pH	SOM	Clay	OF
F1	1									
F2	n	1								
F3	n	n	1							
F4	n	n	0.79**	1						
F5	n	n	n	n	1					
F6	n	n	0.59*	n	0.75**	1				
pH	n	-0.81**	n	n	n	n	1			
SOM	n	-0.63*	0.72**	0.82**	n	0.69*	n	1		
Clay	n	n	n	n	0.64*	0.67*	n	n	1	
TF	n	n	0.64*	n	0.73*	0.99**	n	0.73**	0.64*	1

Table 3/2. Zn

	F1	F2	F3	F4	F5	F6	pH	SOM	Clay	OF
F1	1									
F2	n	1								
F3	n	0.73**	1							
F4	n	n	0.62*	1						
F5	n	n	n	n	1					
F6	n	-0.63*	n	n	-0.72**	1				
pH	n	n	n	0.63*	n	n	1			
SOM	n	n	n	n	n	0.72*	n	1		
Clay	n	-0.65*	-0.61*	n	-0.64*	0.59*	n	n	1	
TF	n	n	n	n	-0.65*	0.99***	n	0.77**	n	1

Table 3/3. Pb

	F1	F2	F3	F4	F5	F6	pH	SOM	Clay	OF
F1	1									
F2	n	1								
F3	n	n	1							
F4	n	0.80**	0.76**	1						
F5	n	0.81**	n	n	1					
F6	n	n	n	n	n	1				
pH	n	n	n	n	-0.77**	n	1			
SOM	n	n	n	n	-0.59*	0.59*	n	1		
Clay	n	n	-0.75*	n	n	n	n	n	1	
TF	n	n	0.67*	n	n	0.86**	n	n	-0.64*	1

Table 3/4. Cd

	F1	F2	F3	F4	F5	F6	pH	SOM	Clay	OF
F1	1									
F2	n	1								
F3	n	n	1							
F4	n	n	n	1						
F5	n	n	0.69*	-0.59*	1					
F6	n	n	n	n	n	1				
pH	n	n	0.77**	n	n	n	1			
SOM	n	n	n	n	n	n	n	1		
Clay	n	n	n	n	n	n	n	n	1	
TF	n	n	0.69*	n	0.67*	0.95***	0.65*	n	n	1

*p < 0.05; ** p < 0.01; *** p < 0.001; n - no correlation

part of the area it follows that organic matter will have a significant role in regulating the behavior of Zn in soil. Much of rainfall and the dry falloff in the town of Sofia are with alkaline character (Gencheva, 1985).

Phosphorus can form insoluble phosphates with HM in urban soils. This could explain the finding that positive correlation between "insoluble residue" (F6), total phosphorus and total contents of Cu and Zn in the urban soils was obtained.

Simple correlation analysis revealed strong positive relationship between Cu_{tot} , F5, F3, organic matter (OM), silt+clay and P contents and in a multiple regression, as well ($p < 0.05$):

$$Cu_{tot} = 4.24 + 3 \times 10^{-4} P + 0.02 (\text{silt+clay } \%) + 0.61 \text{ SOM} \% + 0.99 F6 - 1.1 F5 + 1.4 F3 \quad (R^2 = 0.99).$$

Zn_{tot} was positively correlated with residual forms F6, soil organic matter (SOM) and P and in a multiple regression:

$$Zn_{tot} = 12.85 + 0.02 P + 1.08 \% \text{ SOM} + 0.74 F6 \quad (R^2 = 0.99).$$

Total Pb was correlated with CO_3 -bound Pb, residual Pb F6 and negatively correlated with silt+clay content:

$$Pb_{tot} = 7.32 - 0.02 \% (\text{silt+clay}) + 1.9 F3 + 0.84 F6 \quad (R^2 = 0.93).$$

Labile Pb in F2 was positively correlated with OM-bound Pb (F4) and Fe-oxide bound Pb (F5):

$$Pb_{F2} = -0.005 + 0.19 F4 + 0.32 F5 \quad (R^2 = 0.80).$$

Total Cd was positively correlated with CO_3 -bound F3, Fe-oxide bound and residual F6 forms and in a multiple regression:

$$Cd_{tot} = 0.05 + 1.5 F3 + 0.4 F5 + F6 \quad (R^2 = 0.99).$$

Regression equations demonstrated that the variations in the total pools of metals were explained by the respective contents in F6, P and other variables in the soil. For Cu and Cd, 99 % of the variability in the total Cu and Cd concentrations is explained by a multiple linear regression model with F6, F5, and F3 as predictor variables.

For the mobile PbF2 contents the explained variance reached 80 % based on F4 and F5 fractions. The observed relationships can be due to the fact that total contents were mainly concentrated in the silt+clay, SOM and F6 fractions, while Cu (total) might be occluded in insoluble phosphates, as well.

Total contents ($mg \cdot kg^{-1}$) of Cu (36.3 ± 21.8), Zn (59.2 ± 29.8), Pb (29.1 ± 5.3) and Cd (0.32 ± 0.11) do not exceed limits for agricultural, residential, recreational and institutional sites established in the Bulgarian standards (Regulation No. 3/1979, Regulation No. 3/2008), which are the following (pH < 6 and pH 6-7,4): Cd 1,5 and 2 mg/kg; Cu 80 and 150 mg/kg; Pb 60 and 100 mg/kg; Zn 200 and 320 mg/kg.

A statistically significant ($p < 0.05$) correlation coefficient ($R^2 = 0.65$) between F4 fraction for copper and copper in fine roots was obtained (Figure 2). This dependency shows that in urban soils under oak vegetation, a large part of Cu in F4 might be available and can accumulate in tree roots. The high content of organic matter increases bio-accessibility of Cu, but reduces the accumulation of Pb in fine roots of oak trees. For decreasing bio-accessibility of Pb, Cd and Zn a high level of P in the soil contributed, although the studied urban soils are low in phosphorus (except U4).

Conclusion

The analysis of sequential extractions of heavy metals (Cu, Zn, Pb, Cd) and assessment of various pools contributes significantly to clarifying their behavior in urban soils and their uptake by tree roots. Based on the analysis of the behavior of metals in urban soils under

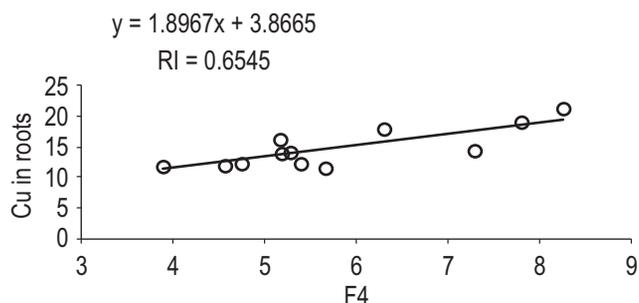


Figure 2. Relationship between Cu in fine roots and "organically bound copper" in soils from urban oak ecosystems of Sofia

oak ecosystems, we found that heavy metals are mainly represented by their residual pools, which are unavailable to plants. With soil organic matter are related mainly Cu and Pb, while with carbonates and Fe and Mn hydroxides are associated mainly Zn and Cd. Mobility and accessibility of the studied heavy metals in urban soils decreases in the order: $Cd > Zn > Cu \geq Pb$. Lead was imported to the soils mainly with dust particles (particulate matter) and accumulated in the surface layers of soil from urban parks. The lowest percentages of soluble and labile forms are determined in the soils with higher pH and Cation exchange capacity measured at pH = 8.2 ($CEC_{8.2}$). Mobile forms of Cu, Zn and Pb are highest in top-soils (0-5 cm), and Cd migrates in depth (0-20cm). Carbonate and organic fractions are presented in small amount. In soils with heavy texture and finer texture (Western park), urbanization influence in relation with the available forms is the least. The data on the chemical fractionation and linear and multiple regression analysis confirmed previous results that Cd and Pb are inferred as markers of anthropogenic pollution, Cu contents are attributed to litho(pedo)genic sources, while Zn is both of anthropogenic and litho(pedo)genic origin. The results obtained for the major city forest parks in Sofia, the largest city and capital of Bulgaria, allow for their use for recreation purposes.

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